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### Natural cultch type influences habitat preference and predation, but not survival, in reef-associated species

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1 **Natural cultch type influences habitat preference and predation, but not survival, in reef-**  
2 **associated species**

3 *Cultch type influences species and interactions*

4  
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18 EAL, JSG conceived and designed the research; EAL, JSG, EMG, CRM performed the  
19 experiments; EAL, JSG analyzed the data; EAL, JSG, EMG, CRM wrote and edited the  
20 manuscript.

21

22 **Abstract**

23 A shared origin with fresh and dredged cultch and availability via mining have made fossil  
24 cultch a commonly used reef restoration substrate. However, important differences in shape and  
25 size between whole-shell cultch and fossil cultch may impact the complexity of reefs constructed  
26 from these materials. To determine if these differences may impact the development of restored  
27 reefs, we quantified the interstitial space each cultch type provides and constructed reef  
28 mesocosms to measure (1) the immediate effects of exposure to each cultch type on mortality of  
29 blue crab (*Callinectes sapidus*) and pink shrimp (*Farfantepenaeus duorarum*); (2) the tendency  
30 of crab, shrimp, and Florida crown conch (*Melongena corona*) to be found on habitats composed  
31 of each substrate type and their position within each in split-substrate mesocosms; and (3) the  
32 influence of cultch type on predation of Eastern oysters (*Crassostrea virginica*) by crabs and  
33 conch. Aggregation of fossil cultch contain more shells and provide less interstitial space than  
34 an equivalent volume of whole-shell cultch. Although immediate mortality following  
35 deployment was low and did not differ among cultch types, we found that all species were more  
36 likely to be found on fresh cultch over fossil cultch in choice experiments and used each habitat  
37 type differently. Cultch type also impacted the size of oysters consumed by crabs in short-term  
38 feeding trials. The structure and traits of habitats created by various materials should be added to  
39 the growing list of issues considered when natural communities are to be restored in oyster reefs  
40 and other environments.

41 **Key words:** *Callinectes sapidus*; *Crassostrea virginica*; *Farfantepenaeus duorarum*; habitat  
42 complexity; interstitial space; *Melongena corona*; oyster reef; restoration

43

44 **Implications:**

- 45 • Differences in the number of shells and total interstitial space per unit volume exist  
46 between whole-shell fresh cultch and broken-shell fossil cultch, with fossil cultch  
47 aggregations packing more tightly and providing less interstitial space.
- 48 • Reef-associated organisms are more likely to be found on fresh cultch in mesocosm  
49 habitats containing both fresh and fossil cultch and use these two cultch habitats  
50 differently. Some predator-prey interactions also differ based on cultch type.
- 51 • In addition to chemical composition, the size and shape of materials used for oyster reef  
52 restoration may affect reef restoration progress.
- 53 • Restoration efforts, and selection and design of materials, should aim to replicate natural  
54 shell shape and variation.
- 55 • Considering reef-associated organisms when designing and monitoring restoration  
56 projects may be crucial to gauging progress and success.

## 57 **Introduction**

58 The materials present in an area and the way those materials are arranged determine the  
59 structure and complexity (e.g., variation in surface area and relief) of a site and can influence  
60 organisms, interactions, and communities (Byrne 2007). For example, organisms may require  
61 specific conditions to persist in the environment, and habitat structure may contribute to meeting  
62 those needs (e.g. refuge space) or to modifying other factors (e.g., hydrodynamic pressure) to be  
63 within a tolerable range. Habitat complexity can also influence species interaction; for example,  
64 predation rates may be altered due to the ability of predators to perceive and pursue prey  
65 (Crowder & Cooper 1982; Penczykowski et al. 2014; Boeye et al. 2014). Because of this, habitat  
66 considerations are critical in both ecological research design and restoration projects. This may

67 be especially important when focal organisms are the ones that actually define communities via  
68 their contribution to structure, such as in mussel beds and kelp forests (Gosnell et al. 2012).

69         The eastern oyster (*Crassostrea virginica*, Gmelin 1791) forms important habitat and has  
70 received a great deal of attention in both research and restoration activities. By forming complex  
71 structures that provide spaces for feeding, nesting, and refuge from predators and other  
72 environmental stressors (Bartol et al. 1999 p. 199; Bartholomew et al. 2000; Gutiérrez et al.  
73 2003; Coen et al. 2007), oysters play a crucial role as ecosystem engineers. The complex habitat  
74 created by oyster reefs supports a suite of species not found in nearby habitats (Coen &  
75 Luckenbach 2000; Tolley & Volety 2005; Scyphers et al. 2011; Brown et al. 2014) and variation  
76 in complexity impacts survival and interactions among species (Lenihan 1999; Grabowski 2004;  
77 Grabowski et al. 2008). Reefs also serve important roles in the protection of shorelines (Coen et  
78 al. 2007; Furlong 2012) and the stabilization of marsh areas (Meyer et al. 1997; Piazza et al.  
79 2005; Scyphers et al. 2011). With the loss of up to 85% of worldwide oyster reefs in the last  
80 century through overharvesting, disease, and dredging (Cohen & Zabin 2009; Beck et al. 2011),  
81 major conservation and economic concerns have emerged. In response to these concerns,  
82 restoration efforts have increased in an attempt to preserve and enhance associated fisheries.

83         Adding fresh cultch (whole oyster shell obtained from shucking houses, restaurants, or  
84 other sources) to degraded reefs was once the primary method of reef restoration. However, the  
85 concurrent decline in oyster populations and increased demand for substrate for use in reef  
86 restoration, along with the loss of oyster shells to landfills, road construction, and the historic use  
87 of tabby concrete, has led to an increase in the use of alternative materials for reef restoration.  
88 Reefs have been constructed from alternative abiotic materials, such as rock, limestone, and  
89 concrete (Furlong 2012; La Peyre et al. 2014), and alternative biotic substrate options, including

90 shells from the surf clam *Spisula solidissima* (Schulte et al. 2009b). A lack of fresh cultch has  
91 also led to extensive efforts focused on procuring additional oyster shell by dredging degraded  
92 reefs (hereafter, *dredged cultch*) or mining ancient oyster beds (hereafter, *fossil cultch*).

93         Although dredging and mining cultch can often have unwanted side effects (e.g.,  
94 disturbance of substrate and organisms) and older cultch may not persist for as long as rock and  
95 concrete, the use of cultch in any form has been encouraged as it may promote reef development  
96 and is considered a more natural source that will not need to be eventually removed from sites  
97 (Schulte et al. 2009b). For example, oyster shell may attract more larvae than artificial substrates  
98 and support different community assemblages, though results from various experiments differ  
99 and outcomes may be location specific (Gibbons et al. 1989; Schulte et al. 2009b; La Peyre et al.  
100 2014; George et al. 2015). Dredged cultch may also last longer than fresh cultch in low-pH  
101 environments (Waldbusser et al. 2011) and potentially prolong reef persistence (Luckenbach et  
102 al. 2005; Waldbusser et al. 2011), allowing more time for the development of a reef community.  
103 The higher availability of dredged relative to fresh cultch has led to its growing use and study  
104 over the past twenty years (Schulte et al. 2009a).

105         Although fossil cultch may share some traits with dredged cultch, it differs markedly in  
106 shape from whole-shell fresh and dredged cultch. Whereas dredged material often still  
107 resembles oyster shells with eroded or rounded edges (for images, see Waldbusser et al. 2011),  
108 fossil cultch more closely resembles gravel, likely due to prolonged time spent weathering over  
109 geologic time scales and the mining process itself (Appendix 1). Reefs are constructed by  
110 depositing several inches of oyster shell or alternative substrate on sites identified as suitable for  
111 reef development or restoration (Schulte et al. 2009b). The smaller, more uniform pieces of  
112 fossil cultch visually appear to form reefs that are more compact and dense compared to reefs

113 constructed of whole-shell cultch due to the wide range in size and shape of fresh cultch pieces.  
114 For these reasons, reefs formed of fossil cultch may differ from reefs formed of whole-shell  
115 natural or dredged cultch in habitat structure, including interstitial space, which is an important  
116 habitat characteristic in determining reef outcomes (Schulte et al. 2009b; Brown et al. 2014).  
117 Interstitial spaces and the habitat heterogeneity they create across tidal gradients are crucial for  
118 recruitment and survival of oysters and benthic macroinvertebrates (Bartol et al. 1999; Soniat et  
119 al. 2004; Tolley & Volety 2005; Gregalis et al. 2009; Brown et al. 2014) and influence local  
120 sedimentation and hydrodynamics (Soniat et al. 2004; Schulte et al. 2009a). The size and  
121 arrangement of cultch pieces might also influence reef stability, movement potential, or vertical  
122 relief. These differences may be crucial to reef health and stability and important to consider in  
123 choosing cultch type for restoration projects. Yet, to our knowledge, such factors have not been  
124 considered, even as fossil cultch has become the primary source of shell material for major  
125 restoration projects (Schulte et al. 2009b; Blankenship 2013).

126         In addition to the noted variations in size and shape, there are additional perceived but  
127 untested differences of impacts to the reef environment between these cultch types. In fact, we  
128 were initially contacted about impacts of fossil cultch by management agencies who had been  
129 contacted with concerns from local fishermen (Apalachicola, FL, USA) about the sediment  
130 released during fossil cultch deployment and its potential impacts on reef-associated species.  
131 The mining and weathering of fossil cultch produces small particles of shell dust that are then  
132 released into the water column, creating a noticeable plume of dust clouding the water until the  
133 particles settle out. This dust was proposed to have negative effects on economically important  
134 reef-associated species. Although most assessments of restorations focus on the presence and  
135 density of oysters or on the number of harvestable oysters (Schulte et al. 2009a; Beck et al. 2011;

136 Furlong 2012; La Peyre et al. 2014), the biological diversity of restored reefs may also be an  
137 important factor in the selection of materials and techniques when restoring reefs and in  
138 engendering the support of local stakeholders.

139         The use of fossil cultch presents an opportunity to investigate how habitat complexity  
140 influences organisms, interactions, and communities. Understanding these differences would  
141 also inform our understanding of trade-offs associated with different substrate use in oyster reef  
142 restoration projects. To assess these concerns, we quantified differences in traits (available  
143 interstitial space per unit volume, number of shell components, average shell volume and weight)  
144 between fresh and fossil cultch and designed a suite of mesocosm experiments to determine  
145 differences between fresh and fossil cultch. These experiments addressed the following  
146 questions: (1) Do the particles released during deployment of each cultch type lead to a  
147 difference in immediate mortality in economically important species located on the site? (2) Are  
148 macroinvertebrates (e.g., blue crab (*Callinectes sapidus*, Rathbun 1896), pink shrimp  
149 (*Farfantepenaeus duorarum*, Burkenroad 1939), and Florida crown conch (*Melongena corona*,  
150 Gmelin 1791)) more likely to be found on reefs composed of either cultch type in mixed reef  
151 mesocosms, and does their habitat usage differ among reef habitats? and (3) Does cultch type  
152 affect predation rates of blue crab and crown conch on Eastern oysters (*C. virginica*)? If cultch  
153 type affects community membership or predation rates, it could influence the speed with which  
154 communities develop and the trajectory they follow, ultimately influencing the ecological and  
155 economic outcomes of restoration projects. If mortality, preference for certain habitats, or  
156 restriction to certain spaces results in species being absent from cultch deposit sites or present at  
157 lower densities, species diversity would be lower at sites using less suitable materials for  
158 restoration than at other restoration sites. The absence of reef-associated species from restoration



159 sites could also reduce support from local stakeholders, including fishermen, for restoration  
160 projects. Together, this set of experiments aimed to better our understanding of how the use of  
161 fresh and fossil cultch might impact the life cycle of reef restoration efforts.

## 162 **Methods**

### 163 Overview

164 All mesocosm experiments were conducted between June and November 2014 in outdoor  
165 mesocosms at the Florida State University Coastal and Marine Laboratory (St. Teresa, FL,  
166 USA). Sixteen tanks (1.2 m diameter, 0.3 m height) were set up on an outdoor platform.  
167 Seawater was pumped from the bay adjacent to the lab into the tanks to a depth of ~0.20 m,  
168 similar to the depth of nearby oyster reefs at low tide, at an average rate of 0.18 L/s. Sixteen  
169 0.012 m (diameter) holes were drilled along the top edge of the pools to allow for drainage while  
170 preventing organisms from escaping. Aluminum flashing wrapped around each of the tanks  
171 prevented organisms from escaping or entering the tanks. Water temperature was recorded using  
172 Thermochron iButtons (Maxim Integrated, model DS1921G-F5#) and compared against local  
173 conditions recorded daily by staff at the FSU Coastal and Marine Laboratory. Over the course of  
174 our study, water temperature ranged from 15.0 to 30.6°C, and salinity ranged from 27.7 to 34.6  
175 ppt.

176 Fresh and fossil cultch from the Eastern oyster, *Crassostrea virginica*, was provided by  
177 the Florida Fish and Wildlife Conservation Commission. Fossil cultch was mined north of  
178 Carrabelle, FL (Gulf Coast Aggregates), where it had been rinsed, stored outside, and exposed to  
179 sun and rain before use, a common practice when using cultch (Cohen & Zabin 2009). Cultch  
180 was reused in multiple experiments. Between each set of experiments and trials, cultch was  
181 removed from the tanks, rinsed, and exposed to sun for at least 4 days. Tanks were rinsed,

182 scrubbed, and dried fully to avoid confounding factors of prior cultch type and organism  
183 presence. Organisms used in experiments were either collected from waters directly adjacent to  
184 the lab or purchased from local suppliers. Analyses were carried out using R (R Core Team  
185 2014) and installed packages including ggplot2 (for graphical analysis and figures, Wickham  
186 2009), lme4 (for generalized linear mixed effect models, Bates 2010), boot (for calculating  
187 inverse logit of model coefficients, Canty & Ripley 2014), car (for testing nested models, 2011),  
188 multcomp (for post-hoc tests, Hothorn et al. 2008), plyr (for database manipulation, Wickham  
189 2011), and optimx (to aid in fitting generalized linear mixed effect models, Nash & Varadhan  
190 2014).

#### 191 Quantifying differences between fresh and fossil cultch

192 To estimate differences in habitat traits between fresh and fossil cultch reefs, we  
193 measured cultch piece density, mass, and available interstitial space per unit volume. We  
194 measured these traits by filling a 3-liter beaker of known weight with fresh or fossil cultch and  
195 weighing it, then adding in water to fill the interstitial space among the cultch pieces and  
196 weighing the beaker a final time. We repeated these measures five times for each cultch type.  
197 To estimate how much settling and configuration may differ among trials using the same shells,  
198 we filled the beaker 4 more times with a single set of cultch we had previously used and  
199 measured variation among the fills. Knowing the volume of the container and weight of water  
200 the beaker held when empty and when filled with cultch allowed us to determine the available  
201 interstitial space per liter. Counting the number of shells added each time allowed us to  
202 determine average mass, volume, and density for each cultch type. Differences among cultch  
203 types were tested using linear models with cultch type as a fixed factor (for the shells that we  
204 measured five times, their average value was used in the final analysis).

205 Impacts of immediate exposure

206           The effect of immediate exposure to the particles released into the water by fresh and  
207 fossil shell during deposition and of direct contact for crabs was examined by measuring  
208 mortality of blue crabs and pink shrimp after cultch addition. Cultch treatments (fresh cultch,  
209 fossil cultch, and a no cultch control) were randomly assigned to 12 tanks, with 4 replicates per  
210 treatment. After the tanks were filled and water was continuously flowing, each treatment tank  
211 received 16.3 kg of cultch to create a layer several centimeters thick covering the bottom of the  
212 tank (Schulte et al. 2009b). Since fossil and fresh cultch are both composed of oyster shell, we  
213 used a common mass between experiments to standardize the amount of shell used for each  
214 cultch type. Within 60 s of cultch addition, each tank received 4 crabs with a mean carapace  
215 width of 115 mm and 10 shrimp. Crabs were placed directly in the tanks, and shrimp were  
216 placed together in a single 18.9 L container with 4 mesh-covered holes allowing for water flow  
217 and protection from predators. The separate container for shrimp in the tanks provided  
218 protection from the crabs while allowing both species to be exposed to any shell particles  
219 released. Crabs and shrimp were collected the previous day through typical commercial  
220 harvesting methods: crab traps in Apalachicola Bay, FL, and shrimp nets in Horseshoe Beach,  
221 FL. Crabs and shrimp were checked and dead individuals removed at 16:30 on the first day (4  
222 hours after experimental setup), at 09:00 and 16:30 for the next 4 days, and at 09:00 on the last  
223 day.

224           The effect of cultch treatment on mortality of shrimp and crabs was analyzed using a  
225 generalized linear mixed-effects model since mortality, the number of organisms remaining alive  
226 (successes) contrasted with the number found dead (failures) during each sampling period, was a  
227 binomial variable, and tank was added as an observation-level random effect to account for

228 potential similarities in mortality within experimental units (Harrison 2014). Correlated random  
229 effects were also fit for each tank to account for the repeated measures aspect of the experiment  
230 (Bates 2010). P-values were obtained by comparing nested models using chi-square tests (Zuur  
231 2009).

### 232 Location among and within reef types

233 We used simple location survey experiments to determine whether organisms exhibited  
234 differential patterns of habitat use or access for the two cultch types. Differences may be due to  
235 preferences for one cultch type over another or to limiting factors (if one cultch type did not  
236 provide easily accessible interstitial space or was unable to be burrowed into); preferences or  
237 limitations would suggest that cultch type could affect an organism's presence on oyster reefs.  
238 Since differences in the size and shape of individual cultch pieces would lead to equal weights of  
239 fossil and fresh cultch resulting in different volumes of shell (and thus reefs of different areas),  
240 we used a consistent volume of cultch to create artificial reef mesocosms in each tank. Two  
241 cultch formations were constructed in each tank by adding 0.019 m<sup>3</sup> of fresh cultch to one half of  
242 the tank and 0.019 m<sup>3</sup> of fossil cultch to the other half with an empty corridor (15.2 cm wide)  
243 between them. Reefs were level and equal in height to avoid any bias due to water depth above  
244 the cultch, and reef orientation differed among tanks to reduce bias due to external  
245 environmental factors. Water flow was started at least 6 hours before organisms were added to  
246 the tanks.

247 We ran separate trials for pink shrimp, crown conch, and blue crab. We ran 5 replicates  
248 of 15 shrimp each; 4 replicates of 10 small (<60 mm shell length) crown conch; 4 replicates of  
249 10 large (>78 mm) crown conch; 5 replicates of 4 small (<105 mm carapace width) blue crabs;  
250 and 5 replicates of 3 large (>115 mm carapace width) blue crabs.

251           Each trial was started by placing individuals in the center of the tank simultaneously and  
252 allowing them to acclimate for ~15 hours before starting observations. At 09:00 and 16:30 for 4  
253 days, the number of individuals on the fresh, fossil, or center portions of the tank was counted  
254 during observations that lasted up to 60 seconds. Since the location of organisms on top of or  
255 within the interstitial space of a reef determines predation risk and resource access, we also  
256 recorded the number of individuals found on top of or within reef habitats of each cultch type.  
257 Individuals on top of cultch included those on the sides of the tank above or near the cultch and  
258 those on shell pieces with less than half of their bodies covered by shell. Individuals were  
259 considered to be within the interstitial space of reef habitats if they were tucked under or into  
260 oyster shells and if half or more of their bodies were covered by shell or below the exterior  
261 surface of the reef (i.e., in a depression within the reef surface). Since shrimp moved around  
262 more than crabs and conchs and could more easily get between individual pieces of shell,  
263 observations for shrimp followed additional guidelines. Tanks with shrimp were observed for  
264 180 seconds at a time, and shrimp had to be located within depressions or tucked into shells for  
265 at least 5 seconds before being considered to be within interstitial reef space. Any shrimp not  
266 found within 180 seconds of observation were counted as within fresh cultch since any shrimp  
267 on top of either reef or within the fossil reef habitat could be found more easily. After the initial  
268 acclimation period, organism location was measured by determining the number of individuals  
269 on top of and within each reef during observation periods.

270           Location data were analyzed using a generalized linear mixed-effects model. Because  
271 both types of cultch occupied the same volume, a null hypothesis was developed that organisms  
272 would be found equally on both substrates. If this was true, the intercept for the model would  
273 not deviate significantly from zero. In addition to considering the intercept, we again included a

274 random effect of sampling time within each tank to account for repeated measures and potential  
275 similarities within each tank. P-values were determined based on models fit by maximum  
276 likelihood. For crabs and conchs, we also considered the impact of organism size on location  
277 and retained this variable in the final model only if it was significant; however, including or  
278 excluding size did not change the overall results regarding relative number found on each cultch  
279 type. For the subset of individuals found on each type of cultch, we used a similar model  
280 structure to consider differences in habitat usage, focusing on the numbers of organisms found  
281 on top of reefs as compared to those within interstitial space. Finally, we used a similar model  
282 structure including cultch type as a factor to compare differences in habitat usage between the  
283 substrates for each organism.

#### 284 Predation

285 Effects of cultch type on predation rate by crab and conch on oysters were also examined.  
286 Four cultch treatments (fresh, fossil, mixture of fresh and fossil, and no cultch as a control) were  
287 randomly applied to 16 tanks. Each tank with cultch had 0.04 m<sup>3</sup> of cultch added: either 0.04 m<sup>3</sup>  
288 fresh, 0.04 m<sup>3</sup> fossil, or 0.02 m<sup>3</sup> of fresh mixed with 0.02 m<sup>3</sup> of fossil. For each trial, we  
289 randomly distributed six ceramic tiles (10.80 cm. x 1.90 cm x 0.64 cm) per tank on top of the  
290 cultch so that oysters were accessible to predators. Two tiles had five small (~20 mm length)  
291 oysters glued to each tile, two had five large (~40 mm length) oysters, and two contained a mix  
292 of large and small oysters. We added 2 tiles of control oysters in small cages to each tank (10  
293 oysters: 5 large and 5 small) to ensure oyster death was due to predation rather than tile  
294 preparation or experiment effects. All were single oysters, set on microcultch by hatchery and  
295 raised in waters adjacent to the bay. For the initial conch predation trials lasting 4 days, we  
296 added 4 conchs with a mean length of 70 mm and mean weight of 49 g to each tank. For crab

297 predation trials lasting 2 days, we added 3 crabs with a mean carapace width of 132 mm to each  
298 tank. After the trial period ended, tiles were removed and oyster consumption was noted by the  
299 presence of empty shells or absence of oysters altogether, indicating that the oyster shell was  
300 removed by a predator. In subsequent trials, we reduced the experimental period by half, with 2  
301 day trials for conchs and 1 day trials for crabs, to consider potential short-term differences  
302 among cultch types. New conchs were collected for the 2 day trials (mean length of 69 mm and  
303 mean weight of 49 g). After a holding period during which crabs were not fed, the crabs from  
304 the 2 day trials were used for the 1 day trials. The crabs were mixed and haphazardly assigned to  
305 pools for the 1 day trials.

306 We analyzed oyster consumption using a generalized linear mixed-effects model. Oyster  
307 consumption was modeled as a binomial variable, and the effect of treatment and oyster size was  
308 considered a fixed effect; interactions among oyster size and cultch type were also considered  
309 but were removed from the model if not significant. Tank effects were again included as an  
310 observation-level random effect to account for overdispersion in the data. When significant  
311 differences existed based on treatment, multiple comparison tests were carried out using the  
312 multcomp package in R (Hothorn et al. 2008) to compare differences among fresh, fossil, and  
313 mixed cultch; family-wise error rates were corrected for using the Holm method (Holm 1979).  
314 Provided p-values were computed using Type III analysis of variance tables.

## 315 **Results**

### 316 Quantifying differences among fresh and fossil cultch

317 There was a significant difference in the number of pieces of fresh (mean: 14.13, range:  
318 11.33 - 16) and fossil (mean: 146.67, range: 138.33 - 153.33) cultch per L ( $t = -46.230, p <$   
319 0.001). Fossil cultch also had a significantly lower average weight per piece than fresh cultch

320 (8.12 g vs 45.81 g,  $t = 12.270$ ,  $p < 0.001$ ). There was significantly more interstitial space per L  
321 (hereafter, *proportion of interstitial space*) for fresh cultch than fossil cultch (0.59 vs 0.41,  $t =$   
322 24.03,  $p < 0.001$ ; Fig. 1), which corresponded to a significantly lower volume per shell for fossil  
323 vs fresh cultch (3.8 mL vs 26.0 mL,  $t = 14.205$ ,  $p < 0.001$ ). Fossil cultch also was denser than  
324 fresh cultch (1.99 kg/L vs 1.71 kg/L,  $t = -8.822$ ,  $p < 0.001$ ). Graphical analysis of repeated  
325 measurements indicated that changes that occurred among trials were minimal relative to the  
326 overall difference between fresh and fossil cultch (Fig. 2); for each cultch type the proportion of  
327 interstitial space varied by less than .02 per fill.

### 328 Immediate exposure

329 Low mortality was noted throughout the experiment, with only 7 crabs (14.6%) and 20  
330 shrimp (16.7%) dying. Mortality occurred in all treatments regardless of cultch type, and there  
331 was no significant difference between the treatments in crab ( $X^2_2 = 0.925$ ,  $p = 0.6297$ ) or shrimp  
332 ( $X^2_2 = 5.110$ ,  $p = 0.078$ ) mortality.

### 333 Location among and within reef types

334 Overall, organisms displayed a significant tendency to be located on sides of tanks  
335 containing fresh cultch as opposed to fossil cultch (all  $p \leq 0.018$ ; Table 1, Fig. 3. Pink shrimp  
336 had the largest discrepancy in likelihood to be observed on the fresh or fossil side of mesocosms.  
337 Species also demonstrated different habitat usage within cultch types (Table 1). Shrimp were  
338 more likely to be within interstitial space of reef habitats composed of fresh cultch than on top of  
339 the same habitat type (Intercept Coefficient (IC) = 2.694,  $p < 0.001$ ; Fig. 4), but more likely to be  
340 on top of fossil cultch than within fossil cultch interstitial space (IC = -3.277,  $p < 0.001$ ). The  
341 difference in habitat usage among cultch types by shrimp was significant ( $X^2_1 = 180.17$ ,  $p <$   
342 0.001). Conchs were more likely to be found on top of each type of cultch than within them



343 (fossil cultch:  $IC = -4.464, p < 0.001$ ; fresh cultch:  $IC = -1.753, p < 0.001$ ; Fig. 4). For fresh  
344 cultch reefs, the difference was even stronger for larger conchs, but both sizes had similar  
345 preferences while using fossil cultch. The difference in habitat usage among cultch types by  
346 conchs was also significant ( $X^2_1 = 66.848, p < 0.001$ ). Crabs were more likely to be found in  
347 than on fresh cultch ( $IC = 0.890, p < 0.018$ ), and in fresh cultch habitats small crabs were even  
348 more likely than large to be found in cultch. Although crabs showed no differences in habitat  
349 usage in fossil cultch reef habitats, there was a significant difference in habitat usage among the  
350 two cultch types ( $X^2_1 = 12.002, p < 0.001$ ).

### 351 Predation

352         There was a significant interaction between cultch type and the size of oysters consumed  
353 during 1 day trials of crab predation (Table 2). Post-hoc tests indicated crabs consumed  
354 significantly more small oysters than large oysters on fossil cultch and in control tanks ( $X^2_1 =$   
355  $4.657, p = 0.031$ ;  $X^2_1 = 13.697, p < 0.001$ ), while on fresh cultch they consumed marginally  
356 more large oysters ( $X^2_1 = 3.205, p = 0.073$ ; Fig. 5). No difference in size of oysters consumed  
357 was observed in mixed cultch treatments. For the 2 day trials, more small oysters than large  
358 oysters were consumed overall, but no differences were observed between cultch treatments.

359         While neither cultch treatment nor oyster size affected predation by conchs after 2 days,  
360 both influenced conch predation after 4 days. Overall, conchs consumed more large oysters than  
361 small oysters, and significant differences in oyster consumption existed among cultch treatments  
362 (Table 2). However, comparison between cultch treatments did not reveal any significant  
363 differences among our planned comparisons (fresh vs. fossil, fresh vs. mixed, and mixed vs.  
364 fossil), indicating the difference likely existed between cultch and control treatments.

### 365 **Discussion**

366 Overall, we found clear differences between cultch types in the average size, weight, and  
367 density per shell and in the amount of interstitial space aggregations of each cultch type. In  
368 exposure experiments, the type of cultch used had no immediate impact on the species examined  
369 as a result of cultch deployment. Although the pool mesocosms are highly simplified versions of  
370 the field environment, the exposure to shell dust in the confined pool may be more concentrated  
371 than it would be in a natural reef setting, thus compounding the potential impacts of shell dust on  
372 mortality. The low mortality observed thus suggests that, even with the more concentrated shell  
373 dust, the cultch type did not impact species mortality. The mortality we observed in blue crabs  
374 and pink shrimp likely resulted from pre-experimental exposure.

375 While cultch type did not affect mortality, blue crabs, pink shrimp, and crown conch  
376 were all more likely to be found on the side of tanks containing fresh cultch as opposed to fossil  
377 cultch. Habitat use on reef habitats composed of each cultch type also differed for all organisms,  
378 with organisms being less likely to be found within reef structures formed of fossil cultch.  
379 Cultch type also influenced how predators consumed oysters, with crabs in short-term trials  
380 consuming more small oysters on fossil cultch but showing no size preference on reefs of mixed  
381 or fresh cultch. Together these results suggest cultch type may impact community structure and  
382 restoration success by influencing the complexity of the habitat and predation interactions  
383 between species.

384 Differences we observed between reefs constructed of these two cultch types may be due  
385 to the noted differences in habitat traits. Our data suggest that reefs composed of fossil cultch  
386 have less interstitial space, and the habitat usage studies indicate this may lead or allow  
387 organisms to use interior reef space less in fossil cultch reefs. Differences in interstitial space are  
388 likely due to the fact that the smaller size of fossil cultch pieces causes them to pack together

389 more tightly, hence the greater number of shells of fossil cultch per liter. This nestled effect  
390 limits the space between fossil cultch pieces. Conversely, the lower density of fresh cultch  
391 pieces may reflect the tendency of these larger, more structurally varied cultch pieces to settle  
392 with more space between the pieces. The complexity of fresh cultch could limit predation on  
393 small organisms by providing more refuge, making prey more difficult to locate, and by being  
394 more difficult for predators to navigate. This may be particularly important for juveniles  
395 recruiting to the reef and may lead to a preference for fresh cultch reefs. As species get larger,  
396 their reliance on available interstitial space decreases, as may their ability to fit into small spaces.  
397 This may explain why large conchs and crabs were found more often on top of fresh cultch than  
398 within the cultch. For these reasons, fresh cultch may be better for building new reefs where  
399 many small organisms arrive via recruitment while fossil cultch may be better as a supplement to  
400 degrading reefs with an already established community.

401         The lack of integration of habitat complexity into studies focused on predator effects and  
402 community ecology in general is a previously cited concern (Baber & Babbitt 2004; Grabowski  
403 et al. 2008; Gosnell et al. 2012) that has often been echoed for management scenarios (Lepori et  
404 al. 2005; Byrne 2007). These results demonstrate that habitat complexity may play an essential  
405 role in determining the trajectory of restored communities by affecting the major species that rely  
406 on the reef and must be considered in addition to the origin of reef materials. They also indicate  
407 that including reef-associated species in the monitoring of restored sites or evaluation of  
408 restoration techniques is essential to determining how closely restored reefs resemble natural  
409 sites. The degree of habitat complexity associated with each cultch type could determine which  
410 drivers affect species more on restored reefs. Fresh cultch, with a greater habitat complexity,  
411 could support a more diverse community by providing more refuge and altering predator-prey

412 interactions (Grabowski 2004; Grabowski & Kimbro 2005; Grabowski et al. 2008), allowing  
413 non-consumptive effects to have more influence. Fossil cultch, which creates less complex  
414 habitat, could limit species abundances and cause communities to be driven more by  
415 consumptive effects, possibly limiting species diversity to fewer large species. Habitat that  
416 promotes consumptive and non-consumptive effects, both of which have been shown to drive  
417 community dynamics on reefs (Grabowski & Kimbro 2005; Hill & Weissburg 2012) and  
418 elsewhere (Preisser et al. 2005), may be essential to rebuilding persistent reef communities that  
419 closely resemble natural sites. Although our results suggest these differences may occur, long-  
420 term monitoring of reefs constructed of varying percentages of fossil and fresh cultch are needed  
421 to fully address these issues. For example, does the use of fossil cultch lead to early changes in  
422 community membership, larval settlement, and interaction strength, such as excluding predators  
423 or changing competitive dynamics, which have long-term repercussions for reef development?  
424 Or does settlement of oysters eventually produce a mixed cultch reef that then progresses to  
425 more closely resemble a naturally occurring reef in species diversity and interactions? On what  
426 time scale do these changes occur, and how do they relate to goals for restoration projects?  
427 Studies of how cultch type influences other interactions (predator-predator for food or space)  
428 may also be important in determining how cultch impacts reef development. Similarly, on a  
429 landscape scale, does using fossil cultch lead to an overall decline in diversity or just cause the  
430 distribution of predators to favor natural cultch sites? Such studies could also offer insight into  
431 how impacts of complexity vary across an organism's life cycle and development. Within a  
432 species, smaller organisms will be able to use interstitial space differently and may be less able  
433 to manipulate cultch, so the effect of cultch type may also differ as organisms grow and change

434 in size. The direct effect of the sediment released by fossil cultch on oysters should also be more  
435 closely considered, as species included here may be less impacted than filter feeders.

436 Our results also suggest that efforts should be made to manipulate abiotic materials to  
437 increase habitat complexity. This may inform the use of current materials such as concrete as  
438 well as new suggested substrates. For example, porcelain is a commonly available material that  
439 is being considered for use as a reef substrate, primarily in crushed form (George 2013).  
440 Varying the size and shape of the crushed material could yield reefs that exhibit complexity  
441 similar to that of natural reefs. These questions should be more directly considered to offer  
442 guidance to restoration efforts.

443 Besides these noted differences in cultch shape, size, and resulting habitat complexity,  
444 other differences may also exist among the cultch types that were not quantified here. Although  
445 all cultch was dried for more than 6 months prior to use, as is common practice for cultch used in  
446 restoration processes, it is possible that biofilms or other biological components may still remain  
447 on fresh cultch, leading to chemical signals that differ among the cultch types. Likewise,  
448 although we have considered macroscale differences in habitat complexity here, weathering and  
449 other processes (including the actual mining) may lead to differences in the texture of cultch that  
450 may impact the ability of larvae to settle or other organisms to move throughout reefs composed  
451 of various materials. These issues warrant further study, especially as they may inform the  
452 actual development of alternative, non-biological cultch replacements. Habitat complexity itself  
453 could also be better measured, as could aspects of habitat components. For example, although we  
454 noted a visible difference in the cloudiness of water following the release of fossil cultch,  
455 sedimentation rates or initial water clarity would offer a more quantitative measure of differences  
456 between cultch type.

457           The relative importance of substrate and complexity may depend on restoration goals and  
458 availability of restoration materials. For example, if the primary goal is to grow oysters or  
459 promote specific ecological services provided by oyster reefs, such as water filtration, without  
460 developing a community, oysters may be housed in floating aquaculture cages. While these are  
461 not traditional reefs, they may be significantly less expensive to produce and may be employed  
462 when other factors, such as substrate availability or the ability to deploy substrate, are limited.  
463 When ecological community health and diversity are the primary goals, using cultch to provide a  
464 more natural substrate as well as greater habitat complexity may be the best choice, and our  
465 results suggest fresh cultch or similarly-shaped materials should be used when possible.

466           However, fresh cultch is in limited supply, and, while most coastal states have shell  
467 recycling programs, oyster shells are not always recycled. Greater incentives, such as the tax  
468 credits of \$1/bushel given in North Carolina (State of North Carolina 2016) and \$5/bushel in  
469 Maryland (Comptroller of Maryland 2016), may help increase shell recycling. Extending these  
470 policies to other states could help to increase the amount of oyster shell recycled, but it still may  
471 not be enough to supply all of the reefs requiring restoration. To extend the use of available  
472 oyster shell, other materials such as clam shell or sand covered by oyster shell (Nestlerode et al.  
473 2007; Schulte et al. 2009b) could be used to supplement fresh cultch and extend its use. Since  
474 fossil cultch is in greater supply than fresh cultch and is chemically more similar to fresh cultch  
475 than are human-made materials, mixing the cultch types together before depositing on reef sites  
476 or covering a base deposit of fossil cultch with fresh cultch could be used to extend the supply of  
477 fresh cultch. Although the latter configuration was not used in these experiments, it may provide  
478 a more effective use of materials by using less of the more limited resource while maintaining  
479 the preferred characteristics of fresh cultch.

480           While further long-term studies comparing fresh cultch, fossil cultch, and mixed fresh  
481 and fossil cultch to human-made materials could increase our understanding of their relative  
482 impacts on oysters reefs, this study has helped to show that a material’s persistence in the  
483 environment and its origin should not be the only factors considered when restoring reefs.  
484 Habitat traits and complexity may also be important in promoting reef development. All three  
485 species we observed utilized habitat differently among reefs composed of fresh and fossil cultch  
486 and were more likely to be found on the structurally complex fresh cultch reefs. Predation rates  
487 on small oysters also differed among cultch types. The higher rates of predation on small oysters  
488 observed on fossil cultch reefs may lead to fewer oysters overall and fewer large oysters over  
489 time, further reducing the potential for habitat complexity to increase at a site and for sites to  
490 contain populations capable of reproducing. Fewer oysters and less refuge space could in turn  
491 cause abundance and diversity of species on less complex reefs to initially be lower than on more  
492 complex reefs and to continue to decline over time despite restoration efforts. If fossil cultch  
493 was used for reef restoration solely because it may persist longer in the environment, the  
494 community attracted to the reef may not be as stable or as healthy as the community associated  
495 with a fresh cultch reef. In order to reach a balance between reef persistence and healthy  
496 community structure, fresh cultch should be used and supplemented with other materials to  
497 extend the number of reefs that can be successfully restored. Overall, future restoration efforts in  
498 oyster reefs and elsewhere should more fully consider how best to create habitat complexity and  
499 promote diversity in order to restore natural communities and achieve other restoration  
500 objectives.

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505 feedback of anonymous reviewers that greatly benefitted this manuscript.



506 **LITERATURE CITED**

- 507 Baber MJ, and Babbitt KJ (2004) Influence of habitat complexity on predator-prey interactions  
508 between the fish (*gambusia holbrooki*) and tadpoles of *hyla squirella* and *gastrophryne*  
509 *carolinensis*. *Copeia* 2004:173–177
- 510 Bartholomew A, Diaz R, and Cicchetti G (2000) New dimensionless indices of structural habitat  
511 complexity: predicted and actual effects on a predator’s foraging success. *Marine*  
512 *Ecology Progress Series* 206:45–58
- 513 Bartol IK, Mann R, and Luckenbach M (1999) Growth and mortality of oysters (*crassostrea*  
514 *virginica*) on constructed intertidal reefs: effects of tidal height and substrate level.  
515 *Journal of Experimental Marine Biology and Ecology* 237:157–184
- 516 Bates DM (2010) Lme4: mixed-effects modeling with r. URL <http://lme4.r-forge.r-project.org/book>. <http://lme4.r-forge.r-project.org/IMMwR/lrgprt.pdf> (accessed 5 August 2013)
- 517
- 518 Beck M, Brumbaugh R, Airoidi L, and Carranza A (2011) Oyster reefs at risk and  
519 recommendations for conservation, restoration, and management. *BioScience* 61:107–  
520 116
- 521 Blankenship K (2013) Maryland getting florida oyster shell to build reef.  
522 [http://www.bayjournal.com/article/maryland\\_getting\\_florida\\_oyster\\_shell\\_to\\_build\\_reef](http://www.bayjournal.com/article/maryland_getting_florida_oyster_shell_to_build_reef)  
523 (accessed 2 February 2016)
- 524 Boeye J, Kubisch A, and Bonte D (2014) Habitat structure mediates spatial segregation and  
525 therefore coexistence. *Landscape Ecology* 29:593–604
- 526 Brown LA, Furlong JN, Brown KM, and La Peyre MK (2014) Oyster reef restoration in the  
527 northern gulf of mexico: effect of artificial substrate and age on nekton and benthic  
528 macroinvertebrate assemblage use. *Restoration Ecology* 22:214–222

529 Byrne LB (2007) Habitat structure: a fundamental concept and framework for urban soil  
530 ecology. *Urban Ecosystems* 10:255–274

531 Canty A, and Ripley B (2014) Boot: bootstrap r (s-plus) functions

532 Coen LD, Brumbaugh RD, Bushek D, Grizzle R, Luckenbach MW, Posey MH, Powers SP, and  
533 Tolley SG (2007) Ecosystem services related to oyster restoration. *Marine Ecology*  
534 *Progress Series* 341:303–307

535 Coen LD, and Luckenbach MW (2000) Developing success criteria and goals for evaluating  
536 oyster reef restoration: ecological function or resource exploitation? *Ecological*  
537 *Engineering* 15:323–343

538 Cohen AN, and Zabin CJ (2009) Oyster shells as vectors for exotic organisms. *Journal of*  
539 *Shellfish Research* 28:163–167

540 Comptroller of Maryland (2016) Oyster shell recycling tax credit.  
541 [http://taxes.marylandtaxes.com/Business\\_Taxes/General\\_Information/Business\\_Tax\\_Cre](http://taxes.marylandtaxes.com/Business_Taxes/General_Information/Business_Tax_Credits/Oyster_Shell_Recycling_Tax_Credit.shtml)  
542 [dits/Oyster\\_Shell\\_Recycling\\_Tax\\_Credit.shtml](http://taxes.marylandtaxes.com/Business_Taxes/General_Information/Business_Tax_Credits/Oyster_Shell_Recycling_Tax_Credit.shtml) (accessed 2 February 2016)

543 Crowder LB, and Cooper WE (1982) Habitat structural complexity and the interaction between  
544 bluegills and their prey. *Ecology* 63:1802–1813

545 Fox J, and Weisburg S (2011) *An r companion to applied regression* Second. Sage.  
546 <http://socserv.socsci.mcmaster.ca/jfox/Books/Companion>

547 Furlong JN (2012) Artificial oyster reefs in the northern gulf of mexico: management, material,  
548 and faunal effects. Master’s thesis, Louisiana State University, Baton Rouge

549 George L (2013) Relative habitat value of alternative substrates used in oyster reef restoration.  
550 Master’s thesis, Texas A&M University, College Station, Texas.  
551 <https://repositories.tdl.org/tamucc-ir/handle/1969.6/492>

552 George LM, De Santiago K, Palmer TA, and Beseres Pollack J (2015) Oyster reef restoration:  
553 effect of alternative substrates on oyster recruitment and nekton habitat use. *Journal of*  
554 *Coastal Conservation* 19:13–22

555 Gibbons MC, Mann R, and Wright LD (1989) Laboratory and field studies of oyster larvae  
556 settlement on three substrates, oyster shell, tire chips, and expanded shale, and the relative  
557 mobility of the three substrates. *Marine Resources Report No 89-3*, Virginia Marine  
558 Resources Commission

559 Gosnell JS, DiPrima JB, and Gaines SD (2012) Habitat complexity impacts persistence and  
560 species interactions in an intertidal whelk. *Marine Biology* 159:2867–2874

561 Grabowski JH (2004) Habitat complexity disrupts predator-prey interactions but not the trophic  
562 cascade on oyster reefs. *Ecology* 85:995–1004

563 Grabowski JH, Hughes AR, and Kimbro DL (2008) Habitat complexity influences cascading  
564 effects of multiple predators. *Ecology* 89:3413–3422

565 Grabowski JH, and Kimbro DL (2005) Predator-avoidance behavior extends trophic cascades to  
566 refuge habitats. *Ecology* 86:1312–1319

567 Gregalis KC, Johnson MW, and Powers SP (2009) Restored oyster reef location and design  
568 affect responses of resident and transient fish, crab, and shellfish species in mobile bay,  
569 alabama. *Transactions of the American Fisheries Society* 138:314–327

570 Gutiérrez JL, Jones CG, Strayer DL, and Iribarne OO (2003) Mollusks as ecosystem engineers:  
571 the role of shell production in aquatic habitats. *Oikos* 101:79–90

572 Harrison XA (2014) Using observation-level random effects to model overdispersion in count  
573 data in ecology and evolution. *PeerJ* 2:e616

574 Hill JM, and Weissburg MJ (2012) Predator biomass determines the magnitude of non-  
575 consumptive effects (nces) in both laboratory and field environments. *Oecologia* 172:79–  
576 91

577 Holm S (1979) A simple sequentially rejective multiple test procedure. *Scandinavian Journal of*  
578 *Statistics* 6:65–70

579 Hothorn T, Bretz F, and Westfall P (2008) Simultaneous inference in general parametric models.  
580 *Biometrical Journal* 50:346–363

581 La Peyre M, Furlong J, Brown LA, Piazza BP, and Brown K (2014) Oyster reef restoration in the  
582 northern gulf of mexico: extent, methods and outcomes. *Ocean & Coastal Management*  
583 89:20–28

584 Lenihan HS (1999) Physical–biological coupling on oyster reefs: how habitat structure  
585 influences individual performance. *Ecological Monographs* 69:251–275

586 Lepori F, Palm D, Brännäs E, and Malmqvist B (2005) Does restoration of structural  
587 heterogeneity in streams enhance fish and macroinvertebrate diversity? *Ecological*  
588 *Applications* 15:2060–2071

589 Luckenbach MW, Coen LD, Ross PG. J, and Stephen JA (2005) Oyster reef habitat restoration:  
590 relationships between oyster abundance and community development based on two  
591 studies in virginia and south carolina. *Journal of Coastal Research*:64–78

592 Meyer DL, Townsend EC, and Thayer GW (1997) Stabilization and erosion control value of  
593 oyster cultch for intertidal marsh. *Restoration Ecology* 5:93–99

594 Nash JC, and Varadhan R (2014) Unifying optimization algorithms to aid software system users:  
595 *optimx for r*. *Journal of Statistical Software* 60:1–14

596 Nestlerode JA, Luckenbach MW, and O’Beirn FX (2007) Settlement and survival of the oyster  
597 *crassostrea virginica* on created oyster reef habitats in chesapeake bay. Restoration  
598 Ecology 15:273–283

599 Penczykowski RM, Hall SR, Civitello DJ, and Duffy MA (2014) Habitat structure and ecological  
600 drivers of disease. Limnology and Oceanography 59:340–348

601 Piazza BP, Banks PD, and La Peyre MK (2005) The potential for created oyster shell reefs as a  
602 sustainable shoreline protection strategy in louisiana. Restoration Ecology 13:499–506

603 Preisser EL, Bolnick DI, and Benard MF (2005) Scared to death? the effects of intimidation and  
604 consumption in predator-prey interactions. Ecology 86:501–509

605 R Core Team (2014) R: a language and environment for statistical computing. R Foundation for  
606 Statistical Computing, Vienna, Austria. <http://www.R-project.org>

607 Schulte DM, Burke RP, and Lipcius RN (2009a) Unprecedented restoration of a native oyster  
608 metapopulation. Science 325:1124–1128

609 Schulte D, Ray G, and Shafer D (2009b) Use of alternative materials for oyster reef construction.  
610 EMRRP Technical Notes Collection (ERDC TN-EMRRP-ER-12), U.S. Army Engineer  
611 Research and Development Center, Vicksburg, MS

612 Scyphers SB, Powers SP, Heck KL, J, and Byron D (2011) Oyster reefs as natural breakwaters  
613 mitigate shoreline loss and facilitate fisheries. PLoS ONE 6:e22396

614 Soniat TM, Finelli CM, and Ruiz JT (2004) Vertical structure and predator refuge mediate oyster  
615 reef development and community dynamics. Journal of Experimental Marine Biology  
616 and Ecology 310:163–182

617 State of North Carolina (2016) Tax credits.  
618 [http://www.dor.state.nc.us/taxes/corporate/generalcredits0708\\_withsupp.pdf#14](http://www.dor.state.nc.us/taxes/corporate/generalcredits0708_withsupp.pdf#14)  
619 (accessed 6 March 2016)

620 Tolley SG, and Volety AK (2005) The role of oysters in habitat use of oyster reefs by resident  
621 fishes and decapod crustaceans. *Journal of Shellfish Research* 24:1007–1012

622 Waldbusser GG, Steenson RA, and Green MA (2011) Oyster shell dissolution rates in estuarine  
623 waters: effects of ph and shell legacy. *Journal of Shellfish Research* 30:659–669

624 Wickham H (2009) *Ggplot2: elegant graphics for data analysis*. Springer New York.  
625 <http://had.co.nz/ggplot2/book>

626 Wickham H (2011) The split-apply-combine strategy for data analysis. *Journal of Statistical*  
627 *Software* 41:1–29

628 Zuur AF (2009) *Mixed effects models and extensions in ecology with r*. Springer, New York  
629

630 **Tables**

631 **Table 1.** Summary of results from models testing whether organisms were more likely to be  
 632 located on the fresh or fossil side of a mixed substrate mesocosm and whether they were more  
 633 likely to be found within the reef for each substrate type. Bolded values indicate significance at  
 634 the  $p = 0.05$  level. IC = Intercept Coefficient. Size only retained in models used to determine  
 635 intercept coefficients for the species when significant.

Species	Location: Fresh cultch			Location: Within fresh			Location: Within fossil		
	IC	Z value	<i>p</i>	IC	Z value	<i>p</i>	IC	Z value	<i>p</i>
Shrimp	2.43	9.321	<b>&lt;0.001</b>	2.694	10.49	<b>&lt;0.001</b>	-3.277	-4.55	<b>&lt;0.001</b>
Conch	0.711	3.478	<b>&lt;0.001</b>	-1.753	-6.509	<b>&lt;0.001</b>	-4.464	-5.05	<b>&lt;0.001</b>
Size	0.141	0.477	0.634	1.245	4.24	<b>&lt;0.001</b>	20.70	0.001	0.999
Crab	0.673	2.36	<b>0.018</b>	-.543	-.818	0.414	-0.159	-0.629	0.529
Size	0.757	1.481	0.138	1.553	1.985	<b>0.047</b>	-0.227	-0.476	0.634

636 **Table 2.** Summary of results from generalized linear mixed-effects models testing the effect of  
 637 cultch treatment and oyster size on conch and crab predation on oysters. For post-hoc tests see  
 638 text. Missing values (--) were not included in the final model. Bolded values indicate  
 639 significance at the  $p = 0.05$  level.

Species	Time Period	Impact of:		
		Cultch Type	Size	Cultch Type & Size
		$X^2_3$ statistic ( $p$ value)	$X^2_1$ statistic ( $p$ value)	$X^2_3$ statistic ( $p$ value)
Crab	1 Day	4.981 (0.173)	<b>4.858 (0.028)</b>	<b>20.254 (&lt; 0.001)</b>
	2 Days	3.261 (0.353)	<b>4.254 (0.039)</b>	--
Conch	2 Days	2.741 (0.433)	0.866 (0.352)	--
	4 Days	<b>8.09 (0.044)</b>	<b>14.630 (&lt; 0.001)</b>	--



640 **Figure Legends**

641 **Figure 1.** Difference in proportion of interstitial space found in fresh and fossil cultch

642 aggregations. Points are arranged to prevent overlay.

643 **Figure 2.** Variation in measurements of proportion of interstitial space found in fresh and fossil

644 cultch aggregations when a single set of shells were placed in beaker five times for each type of

645 cultch. Points are arranged to prevent overlay.

646 **Figure 3.** Location probability of (A) pink shrimp, (B) crown conch, and (C) blue crab

647 determined by number of individuals observed on fossil or fresh cultch sides of a mixed substrate

648 mesocosm. Black dashed lines at 0.50 represent the null hypothesis of no preference.

649 **Figure 4.** Location probability of pink shrimp, crown conch, and blue crab for being in or on

650 fresh (A, C, E) and fossil (B, D, F) cultch determined by the number of individuals observed in

651 each cultch type. Black dashed line at 0.50 represents the null hypothesis of no preference.

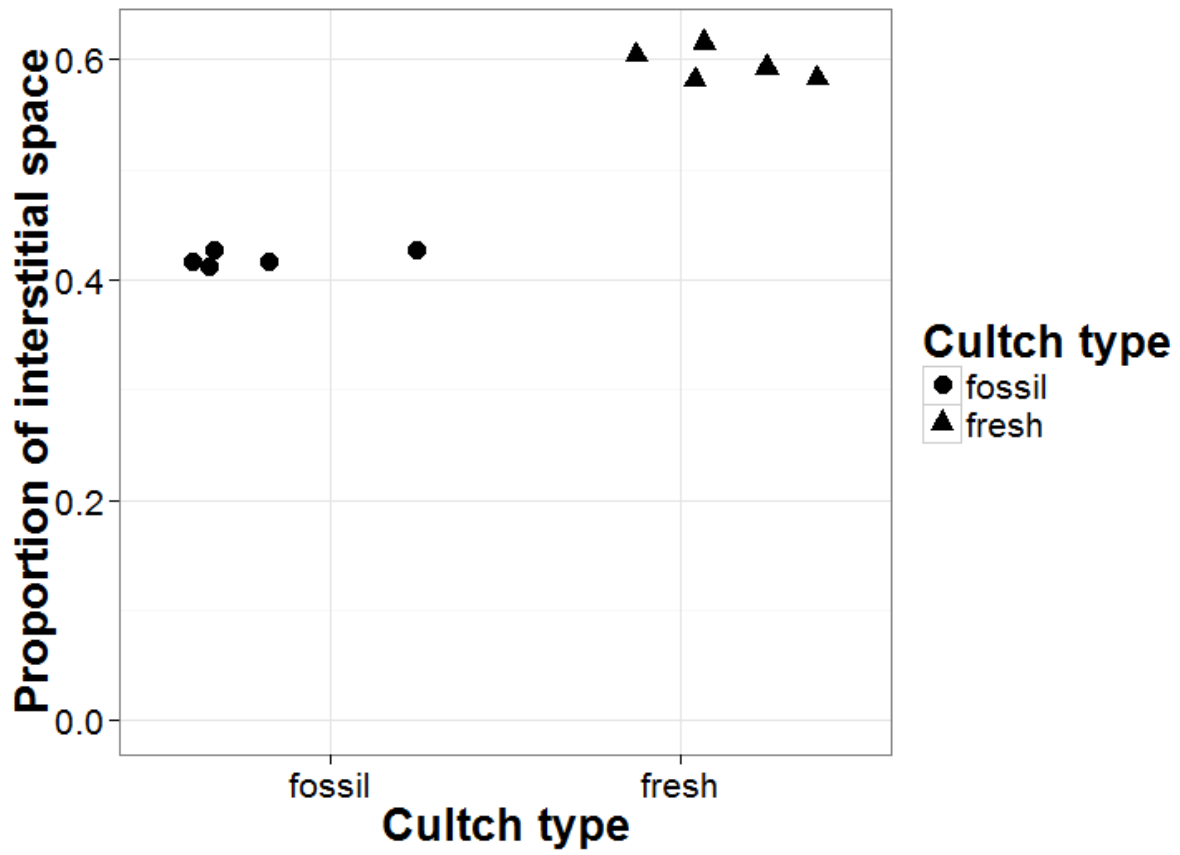
652 **Figure 5.** Predation of blue crab and crown conch upon small (~20mm length, light gray bars)

653 and large (~40mm, dark gray bars) Eastern oysters on fossil, fresh, mixed, and no cultch

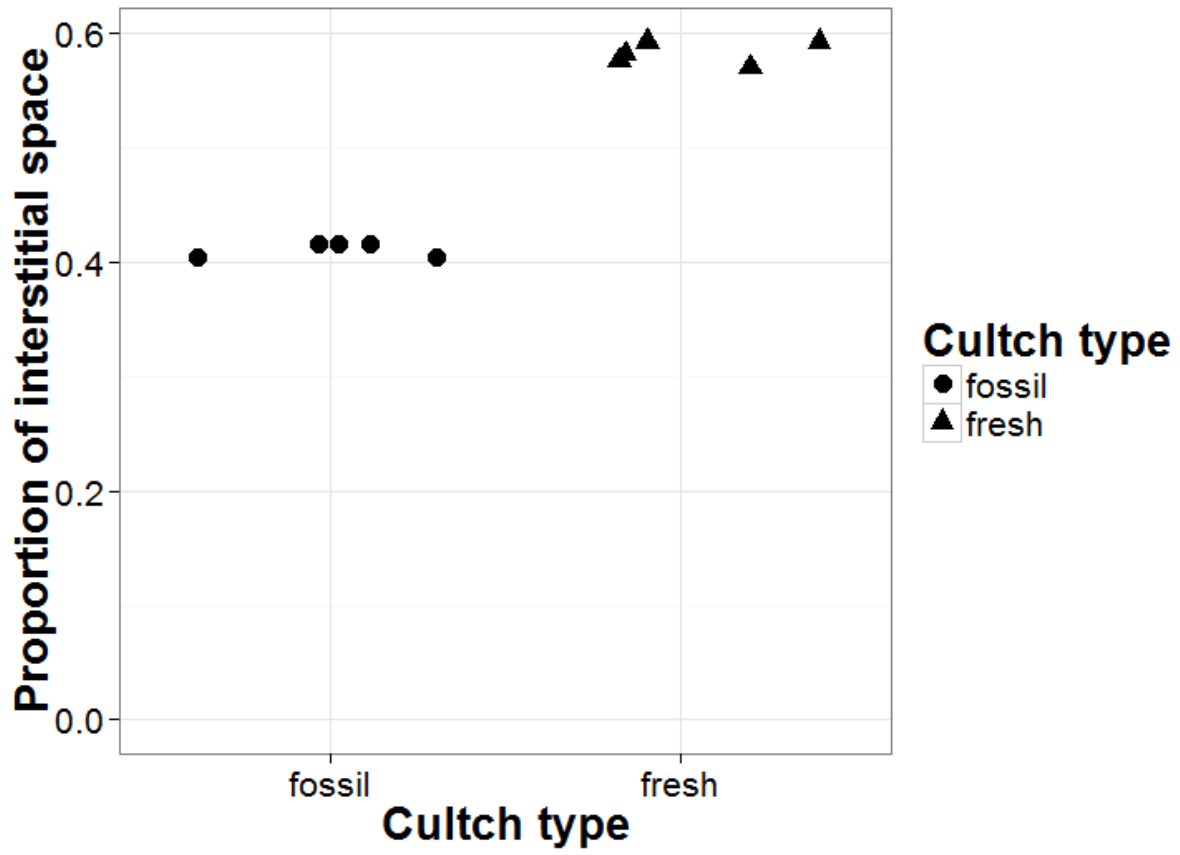
654 treatments with +95% confidence intervals. A) 1-day crab trials, B) 2-day crab trials, C) 2-day

655 conch trials, D) 4-day conch trials.

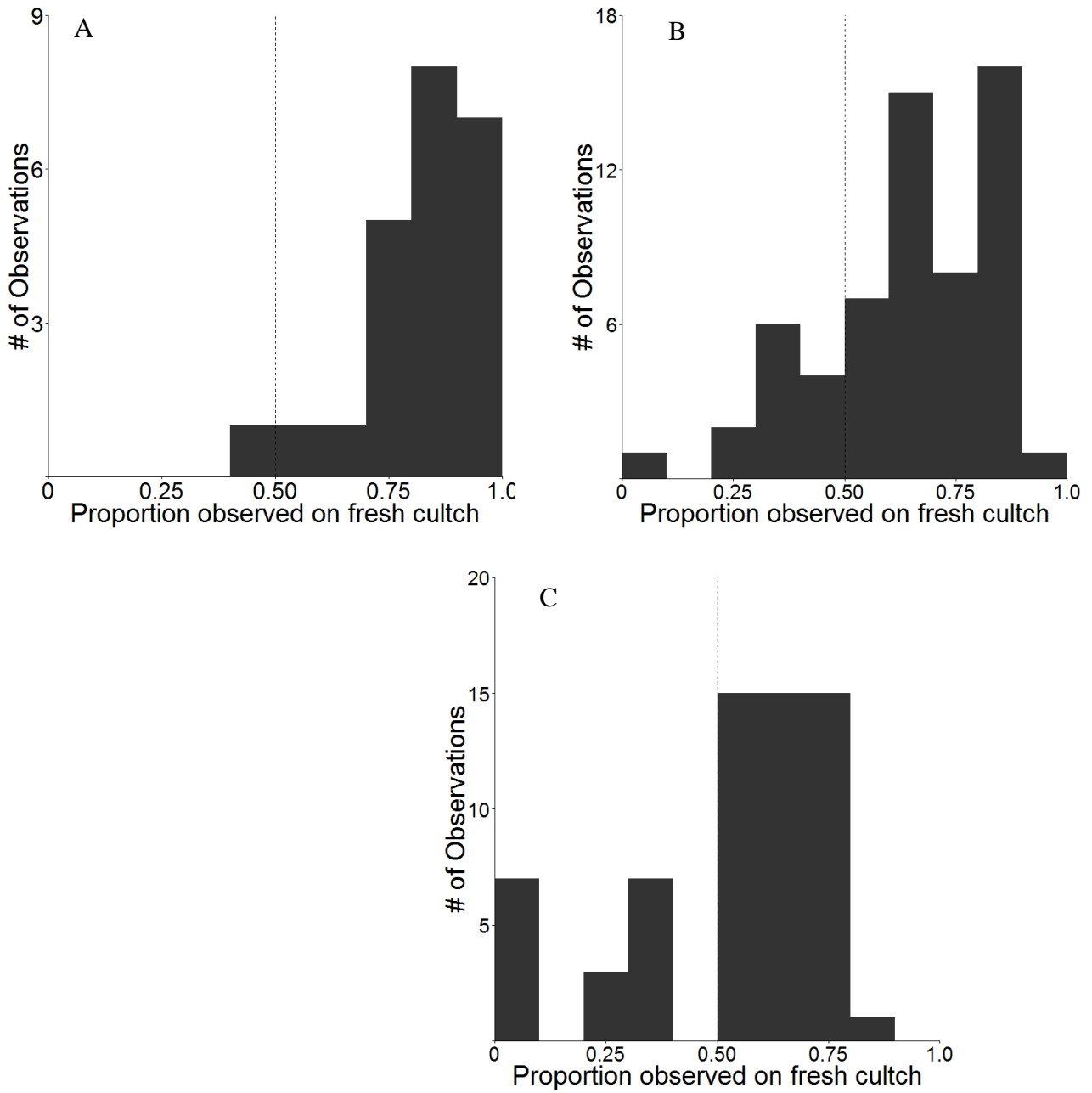
656 Figure 1



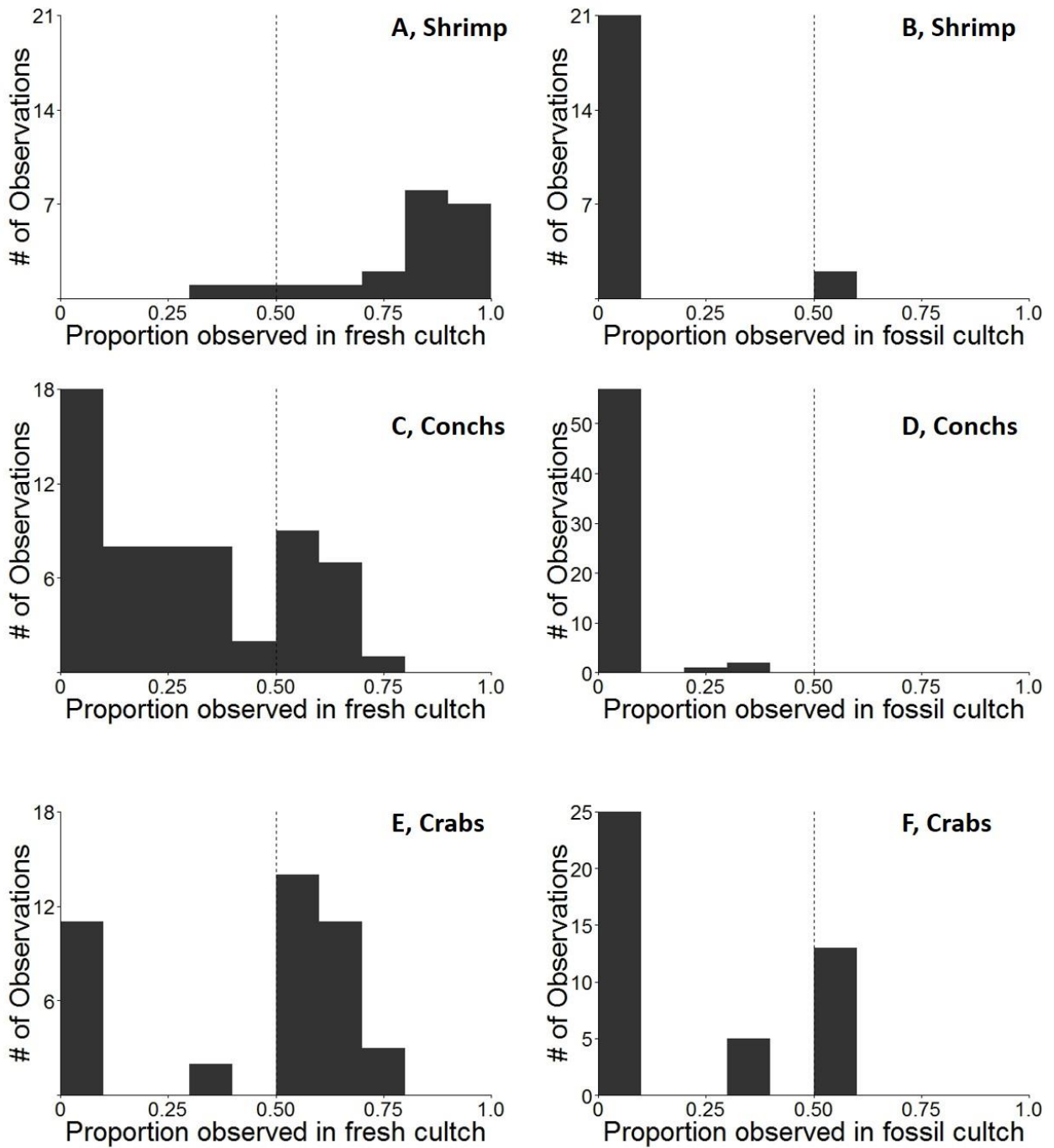
657 Figure 2



658 **Figure 3.**



659 **Figure 4.**



660

